

Effects of two synthetic pyrethroids on *Arthrospira platensis* Gomont growth and antioxidant parameters

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Abstract – The transport of pesticides from application areas to other areas results in pesticide contamination and this sort of contamination has led to unexpected environmental problems worldwide. It is important to determine the responses of phytoplanktonic organisms to these chemicals for an understanding of the effects of pesticides on aquatic ecosystems. In this study, *Arthrospira platensis* Gomont cyanobacteria were exposed to different concentrations of the pesticides cypermethrin (Cyp; 0–50 µg mL⁻¹) and deltamethrin (Dlm; 0–2 µg mL⁻¹). Changes in chlorophyll-*a* concentration, the absorbance at OD560, antioxidant parameters (superoxide dismutase – SOD, ascorbate peroxidase – APX, glutathione reductase – GR, malondialdehyde – MDA, H₂O₂, and proline) were determined under the pesticide exposure. Our results showed that there is a decrease in OD560 absorbance and chlorophyll-*a* content proportional to the increase of pesticide levels. SOD activity decreased with Cyp and Dlm application in *A. platensis* cultures. GR activity also decreased with Cyp applications but did not change with Dlm application. APX activity increased with Cyp treatments but did not change with Dlm applications. Although MDA and hydrogen peroxide contents did not change with Cyp applications, they increased with Dlm applications. Proline contents increased with Cyp applications but decreased with Dlm applications. In conclusion, deltamethrin was more toxic than cypermethrin in the concentrations applied.

Keywords: antioxidant enzymes, *Arthrospira platensis*, cypermethrin, deltamethrin, growth

Introduction

Many chemicals, including pesticides, are released through modern industrial and agricultural activities and have reached a high level in the environment (Burkiewicz et al. 2005). Pesticides affect not only target organisms but also non-target organisms in the aquatic biota (Tremolada et al. 2004). These organisms play an important role in biological processes such as biogeochemical cycling, production, separation, and interaction with other organisms. Pesticides disturb the balance of water ecosystems with their direct action on plants and animals or with their bioaccumulation and transfer abilities in the food chain (Netrawali and Gandhi 1990, Burkiewicz et al. 2005). Planktonic algae display a fundamental role as primary producers (Burkiewicz et al. 2005) and a decrease in algal density and species composition affects the aquatic ecosystem directly by reducing biodiversity and primary production (Li et al. 2005). They are sensitive indicators that allow testing of the different

effects of chemicals released into the water (Burkiewicz et al. 2005). For this reason, microalgae are frequently used in various bioassays (Li et al. 2005). Agricultural chemicals inhibit the growth rate, biomass, and pigment content of freshwater algae by contaminating surface waters in agricultural areas (Netrawali and Gandhi 1990).

Pyrethroids are a class of synthetic insecticides designed and optimized based on pyrethrin structure (Elliott 1995). These pesticides are effective insecticides that are widely used to control agricultural and healthcare pests. After use, they are released into the environment and enter water resources (Mittal et al. 1994). Cypermethrin (Cyp) and deltamethrin (Dlm) have increased toxicity, especially to aquatic organisms with increasing life expectancy (Johri et al. 1997). Megharaj et al. (1987) observed that Cyp has inhibitory effects on *Scenedesmus bijugatus* Kützing. Xiong et al. (2002) and Li et al. (2005) found similar results for *Scenedesmus*

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obliquus Kützing. Wang et al. (2011) carried out a growth inhibition test on *Skeletonema costatum* Cleve, *Scrippsiella trochoidea* (F.Stein) A.R.Loeblich III and *Chattonella marina* (Subrahmanyam) Y.Hara et M.Chihara during 96 h. Sáenz et al. (2012) studied Cyp toxicity to *Scenedesmus quadricauda* Chodat, *Scenedesmus acutus* Meyen, *Chlorella vulgaris* Beyerinck (Beijerinck) and *Pseudokirchneriella subcapitata* (Korshikov) F.Hindák. Wang et al. (2012) tested Cyp effects on *Scenedesmus obliquus*. These references show that Cyp has toxic effects on various groups of microalgae. There are also some studies on Dlm toxicity to aquatic microorganisms. Baeza-Squiban et al. (1987) have shown that the growth of *Dunaliella* sp. Teodoresco and *Chlamydomonas* sp. Ehrenberg was inhibited by Dlm application. Caquet et al. (1992) observed Dlm effects on phytoplankton communities in freshwater mesocosms and found Dlm to disappear rapidly from the aquatic ecosystem.

Antioxidants are compounds that reduce the harmful effects of oxidation via inhibiting free oxygen formation or eliminating free radicals (Baublis et al. 2000). Superoxide dismutase (SOD; EC 1.15.1.1) is a class of metalloprotein that catalyses superoxide to oxygen and hydrogen peroxide (H_2O_2) (Valentine et al. 1998). Ascorbate peroxidase (APX; EC 1.11.1.11) converts hydrogen peroxide to water by using ascorbate as an electron donor and thereby accumulations of H_2O_2 toxic levels are prevented in photosynthetic organisms (Chew et al. 2003). Glutathione reductase (GR; EC 1.6.4.2), is a member of NADPH-dependent oxidoreductases, found in both prokaryotic and eukaryotic cells. GR catalyses the reduction of oxidized glutathione (GSSG) to reduced glutathione (GSH) together with the oxidation of NADPH and plays an important role in the cellular defence systems against reactive oxygen metabolites by creating a reduced GSH pool (Anjum et al. 2010). Malondialdehyde is a metabolite that occurs with peroxidation of lipids including three or more double bonds. Lipid peroxidation causes malondialdehyde formation (Goel and Sheoran 2003). It is known that proline content increases under various stress factors and proline accumulation protects cell parts and cell contents such as enzymes, membranes and polyribosomes (Kishor et al. 2005).

The enzymatic and non-enzymatic antioxidant defence systems create the response to oxidative stress and reflect the tolerance and susceptibility of algae to pesticide exposure. However, although there is some information about algal antioxidant systems, which are either a response to the environmental conditions or a defensive system (Mallick and Mohn 2000) studies about the harmful or activator effects of pyrethroids on algal antioxidant systems are limited.

The aims of our study are: (i) to determine metabolic damage due to cypermethrin and deltamethrin toxicity and (ii) to measure the intracellular level of antioxidant responses for pesticide detoxification in *Arthrospira platensis*-M2 strain. For this purpose, the alterations of some parameters such as growth rate (OD560), chlorophyll-*a* content, superoxide dismutase (SOD), ascorbate peroxidase (APX), glutathione reductase (GR), malondialdehyde (MDA), hydrogen peroxide (H_2O_2) and proline have been investigated during

pesticide application. Responses given by phototrophic microorganisms to pesticides will help us to have an idea of how the communities are affected by the pesticides in the water ecosystems.

Material and methods

Algae culture and treatment

Arthrospira platensis, strain M2, was obtained from the Soley Microalgae Institute (California, USA) (Culture collection No: SLSP01). Algae were grown in Spirulina Medium (Aiba and Ogawa 1977) under axenic conditions. 20 mL algal cultures were inoculated onto 180 mL culture medium in an Erlenmeyer flask and were allowed to grow under full-spectrum lamps providing $93 \mu\text{mol photons m}^{-2} \text{s}^{-1}$ photosynthetically available radiation in a 12:12 h light/dark cycle at $30 \pm 1^\circ\text{C}$ during 10 days. At the end of 10 days, cultures were renewed and, all the flasks contained 50 mL algal culture. Stock solutions of the commercial formulation of Cyp and Dlm (200 g L^{-1} and 25 g L^{-1} , respectively; EC, Sakarya, Turkey) were prepared with distilled water and were then diluted for all bioassays. Various final concentrations of cypermethrin ($10, 20, 30, 40, 50 \mu\text{g mL}^{-1}$) and deltamethrin ($0.125, 0.25, 0.5, 1, 1.5, 2 \mu\text{g mL}^{-1}$) were in the culture medium. The range of concentrations was determined with preliminary range-finding bioassays according to EC50 value for growth parameters and IC50 value for enzyme activity assays and at least five concentrations were selected under these values.

Cell growth and chlorophyll-*a* assay

Optical densities (ODs) at 560 nm were measured spectrophotometrically over 7 days under control and stressed conditions. Chlorophyll-*a* content was estimated by methanol extraction and measured spectrophotometrically over 7 days (MacKinney 1941).

Antioxidant enzyme activities

On the 7th day of the study, 2 mL culture solutions from the control and treated samples were centrifuged at 14,000 rpm for 20 min at 4°C and the resulting pellets were kept at -20°C until enzyme activity was measured. Pellets were ground with liquid nitrogen and suspended in specific buffers with proper pH values for each enzyme. The protein concentrations of algal cell extracts were determined according to Bradford (1976), using bovine serum albumin (BSA) as a standard.

The superoxide dismutase activity was determined by the method of Beyer and Fridovich (1987), based on the photoreduction of NBT (nitroblue tetrazolium). Extraction of pellets (0.2 g) was performed in 1.5 mL homogenization buffer containing 100 mM K_2HPO_4 buffer (pH 7.0), 2% PVP, and 1 mM Na_2EDTA . After centrifugation at 14,000 rpm for 20 min at 4°C , the resulting supernatants were used to measure SOD activity. The reaction mixture consisted of 100 mM K_2HPO_4 buffer (pH 7.8) containing 9.9×10^{-3} M methionine, 5.7×10^{-5} M NBT, 1% triton X-100, and enzyme

extract. The reaction was started by the addition of $0.9 \mu\text{M}$ riboflavin and the mixture was exposed to light with an intensity of $375 \mu\text{mol m}^{-2} \text{s}^{-1}$. After 15 min, the reaction was stopped by switching off the light, and absorbance was read at 560 nm. The SOD activity was calculated by a standard graphic and expressed as $\text{U mg}^{-1} \text{protein}$.

The ascorbate peroxidase activity was determined according to Wang et al. (1991) by estimating the decreasing rate of ascorbate oxidation at 290 nm. APX extraction was performed in 50 mM Tris-HCl (pH 7.2), 2% PVP, 1 mM Na_2EDTA , and 2 mM ascorbate. The reaction mixture consisted of 50 mM K_2HPO_4 buffer (pH 6.6), 2.5 mM ascorbate, 10 mM H_2O_2 , and enzyme-containing 100 μg protein in a final volume of 1 mL. The enzyme activity was calculated from the initial rate of the reaction using the extinction coefficient of ascorbate ($E = 2.8 \text{ mM cm}^{-1}$ at 290 nm).

The glutathione reductase activity was measured with the method of Sgherri et al. (1994). Extraction was performed in 1.5 mL of suspension solution containing 100 mM K_2HPO_4 buffer (pH 7.0), 1 mM Na_2EDTA , and 2% PVP. The reaction mixture (total volume of 1 mL) contained 100 mM K_2HPO_4 buffer (pH 7.8), 2 mM Na_2EDTA , 0.5 mM oxidised glutathione (GSSG), 0.2 mM NADPH and enzyme extract containing 100 μg protein. The decrease in absorbance at 340 nm was recorded. Correction was made for the non-enzymatic oxidation of NADPH by recording the decrease at 340 nm without adding GSSG to assay mixture. The enzyme activity was calculated from the initial rate of the reaction after subtracting the non-enzymatic oxidation using the extinction coefficient of NADPH ($E = 6.2 \text{ mM cm}^{-1}$ at 340 nm).

Determination of malondialdehyde and hydrogen peroxide

The malondialdehyde content was determined by the method of Heath and Packer (1968). 0.2 g of pellet was homogenized in 3 mL of 0.1% TCA (4°C) and centrifuged at 4100 rpm for 15 min and the supernatant was used in the subsequent determination. 0.5 mL of 0.1 M Tris-HCl pH 7.6 and 1 mL of TCA-TBA-HCl reagent (15% w/v) (trichloroacetic acid–0.375% w/v thiobarbituric acid–0.25 N hydrochloric acid) were added to the 0.5 mL of the supernatant. The mixture was heated at 95°C for 30 min and then quickly cooled in an ice bath. To remove suspended turbidity, the mixture was centrifuged at 4100 rpm for 15 min, then the absorbance of the supernatant at 532 nm was recorded. Non-specific absorbance at 600 nm was measured and subtracted from the readings recorded at 532 nm. The MDA content was calculated using its extinction coefficient of $155 \text{ mM}^{-1} \text{ cm}^{-1}$. For determination of the hydrogen peroxide content, 0.5 mL of 0.1 M Tris-HCl (pH 7.6) and 1 mL of 1 M KI were added to 0.5 mL of supernatant. After 90 min, the absorbance was recorded at 390 nm.

The proline content determination

The proline content was determined by the method of Weimberg et al. (1982) where 0.1 g of pellet was homoge-

nized in 10 ml of 3% aqueous sulphosalicylic acid and homogenates were incubated in a hot water bath at 95°C for 30 minutes. The samples were cooled and centrifuged at 4,100 rpm for 10 min. Two mL of the extract reacted with 2 mL of acid-ninhydrine and 2 mL of glacial acetic acid for 1 h at 100°C . The reaction mixture was extracted with 4 mL toluene. The chromophore containing toluene was separated and the absorbance was recorded at 520 nm.

Statistical analysis

The differences between the means of control and treated samples were analysed by one-way analysis of variance (ANOVA), taking $P < 0.05$ as significant according to LSD with two degrees of freedom. Three replicate cultures were used for each treatment. The mean values \pm standard error (SE) are shown in Figures.

Results

Biomass and chlorophyll-*a* content

Biomass production and chlorophyll-*a* content of *Arthrospira platensis*-M2 cells were significantly decreased by Cyp and Dlm application to the culture medium for 7 days ($P < 0.05$, Figs. 1, 2). The most significant reductions were observed at the highest pesticide treatments, ($50 \mu\text{g mL}^{-1}$ for Cyp and $2 \mu\text{g mL}^{-1}$ for Dlm) in biomass production and

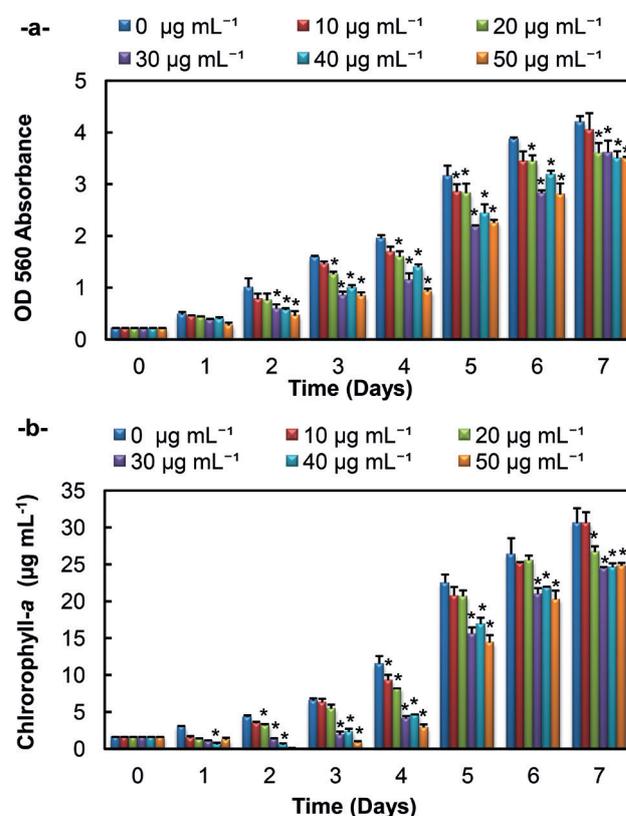


Fig. 1. Biomass values, as absorbance at 560 nm (a), and chlorophyll-*a* concentration (b) of *Arthrospira platensis* treated with 0–50 $\mu\text{g mL}^{-1}$ cypermethrin during 7 days. Data are the means \pm SE of three replicates. *Significantly different from control, $P < 0.05$ (LSD analysis).

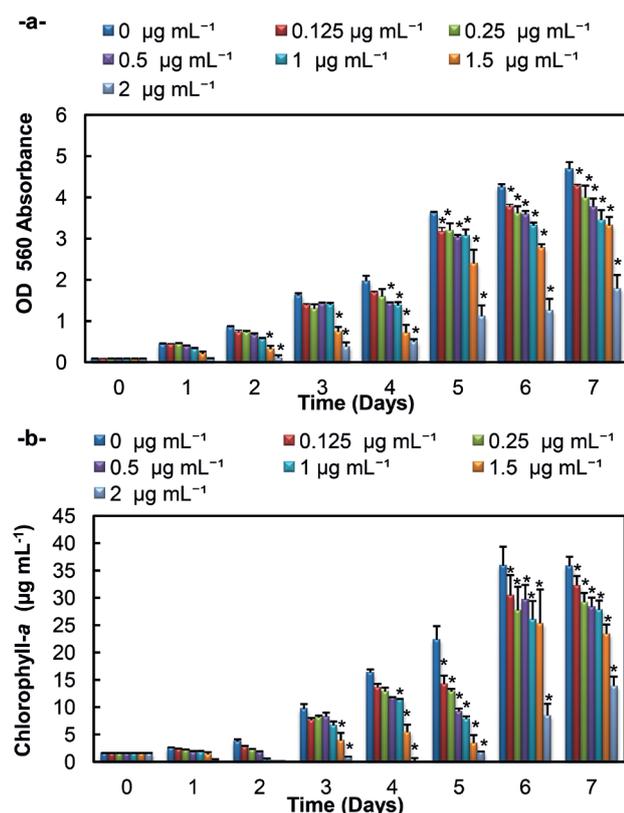


Fig. 2. Biomass values, as absorbance at 560 nm (a), and chlorophyll-*a* concentration (b) of *Arthrospira platensis* treated with 0–2 µg mL⁻¹ deltamethrin during 7 days. Data are the means ± SE of three replicates. *Significantly different from control, $P < 0.05$ (LSD analysis).

chlorophyll-*a* content. Cyp and Dlm toxicity affected the biomass production and chlorophyll-*a* content of *A. platensis*-M2 cells, which increased in a time-dependent manner during the 7 days (Figs. 1a,b, 2a,b). The EC₅₀ values of Cyp and Dlm applications were 53.12 ± 1.16 µg mL⁻¹ and 1.76 ± 0.07 µg mL⁻¹ respectively for 7th day, based on biomass production.

The antioxidant parameters

The effect of Cyp applications on SOD activity in *A. platensis*-M2 cells is presented in Fig. 3a. All Cyp concentrations (5, 10, 20, 30, 40, 50 µg mL⁻¹) caused an increase in SOD activity on *A. platensis*-M2 cells as compared to control. The maximum inhibition in SOD activity (96%) occurred at 40 µg mL⁻¹ Cyp concentration. APX activities increased at the same concentrations ($P < 0.05$, Fig. 3b). On the other hand GR activities also decreased at all concentrations except 5 µg mL⁻¹ Cyp application ($P < 0.05$, Fig. 3c). Ten, 20, 30, 40, 50 µg mL⁻¹ Cyp concentrations caused 17, 66, 47, 70, 85% decrease in GR activity, respectively. MDA and H₂O₂ contents (Fig. 4a,b) did not change significantly, but proline contents (Fig. 4c) increased at all Cyp concentrations ($P > 0.05$).

The effect of Dlm applications on SOD activity in *A. platensis*-M2 cells is presented in Fig. 5a. SOD activities decreased by 37, 47, 21, 39, 36, 52% ($P < 0.05$) at all Dlm con-

centrations, respectively; APX and GR activities did not display significant alteration ($P > 0.05$, Fig. 5b,c). MDA contents increased at 0.5, 1, 1.5 µg mL⁻¹ Dlm applications 74, 85, 78%; respectively ($P < 0.05$, Fig. 6a); however H₂O₂ content increased with 1, 1.5 and 2 µg mL⁻¹ Dlm applications ($P < 0.05$, Fig. 6b). Proline content decreased significantly at 2 µg mL⁻¹ Dlm concentration ($P < 0.05$, Fig. 6c).

Discussion

In this study, we have found that exogenous addition of different concentrations of two synthetic pyrethroids (cypermethrin and deltamethrin) showed varying degrees of toxicity to the growth of *Arthrospira platensis* M2 strain. In a study of *Scenedesmus bijugatus*, *Nostoc linckia* Bornet ex Bornet et Flahault, *Synechococcus elongatus* (Nägeli) Nägeli and *Phormidium tenue* Gomont, Megharaj et al. (1987) found that cypermethrin inhibited the growth of *Scenedesmus bijugatus* at 10–50 µg mL⁻¹ concentrations. Xiong et al. (2002) applied cypermethrin to *Scenedesmus obliquus* during a 96 h period and found that the EC₅₀ was 112 mg L⁻¹. Li et al. (2005) reported that cypermethrin inhibited the growth of *Scenedesmus obliquus* at 50–250 mg L⁻¹ concentrations during 96 h, and the chlorophyll-*a* and carotenoid content of *Scenedesmus obliquus* decreased with the application. Li et al. found the EC₅₀ value of cypermethrin as 112 ± 9 mg L⁻¹. Wang et al. (2011) carried out a growth inhibition test about cypermethrin toxicity on *Skeletonema costatum*, *Scrippsiella trochoidea*, and *Chattonella marina* during 96 h and they determined EC₅₀ values as 71.4 µg L⁻¹, 205 µg L⁻¹ and 191 µg L⁻¹, respectively. Sáenz et al. (2012) study cypermethrin toxicity to *Scenedesmus quadricauda*, *Scenedesmus acutus*, *Chlorella vulgaris* and *Pseudokirchneriella subcapitata* and the inhibited growth concentrations are 0.3–5 mg L⁻¹, 0.6 mg L⁻¹, 0.15–2.5 mg L⁻¹ and 0.075 mg L⁻¹, respectively. In another study, Wang et al. (2012) determine the toxic effects of commercial cypermethrin on *S. obliquus* and measure EC₅₀ value of 2.37 mg L⁻¹. The data and the concentrations used (10–50 µg mL⁻¹) of Megharaj et al. (1987) are the closest to the results of our study. As was shown in the literature, Cyp has an inhibitory effect on algal growth and chlorophyll-*a* content.

As with the toxicity of Cyp, there are some studies about the effect of Dlm on algae. In the study of Burkievicz et al. (2005), the cell density of *Scenedesmus subspicatus* Chodat decreased with 2.5, 5 and 10 mg L⁻¹ Dlm applications during 24 h and 1.25 mg L⁻¹ Dlm application during 72 h. Lutnicka et al. (2014) observed that Dlm has inhibited the growth of *Chlorella vulgaris* by 13% at the end of 14 days with a 0.02 µg L⁻¹ Dlm application.

Although the Dlm concentrations in the study of Lutnicka et al. (2014) were lower than in our applications, the concentrations of Burkievicz et al. (2005) were similar to those in our studies.

In our study, SOD activities of *A. platensis* decreased at all Cyp applications except 5 µg mL⁻¹ and at all Dlm appli-

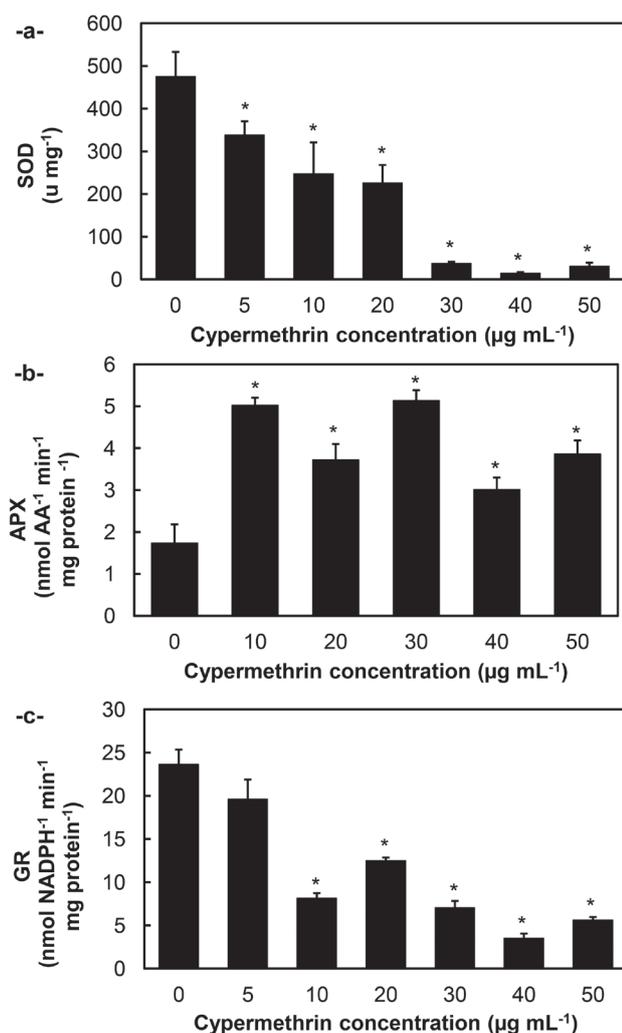


Fig. 3. Total superoxide dismutase (SOD; a), ascorbate peroxidase (APX; b) and glutathione reductase (GR; c) activities of *Arthrospira platensis* treated with 0–50 µg mL⁻¹ cypermethrin. Data are the means ± SE of three replicates. *Significantly different from control, $P < 0.05$ (LSD analysis).

cations according to control. In our study, Cyp and Dlm pesticides also caused significant reductions in chlorophyll-*a*, and thus loss of photosynthetic metabolism may have caused significant reductions in SOD activity. Wang et al. (2011) also reported that Cyp pesticide in high concentrations (> 50 µg L⁻¹) inhibited SOD activity in *Skeletonema costatum*, *Scrippsiella trochoidea* and *Chattonella marina*, and they pointed out the inactivation of SOD activity may have been due to growth being inhibited by cypermethrin. One or more of the suggested reasons in these articles may be the main mechanism of Cyp and Dlm toxicity on the SOD enzyme activity in our study.

GR and glutathione are effective on the Halliwell / Asada pathway in the inactivation of H₂O₂ in plant cells (Bray et al. 2000). GR catalyzes the last step of the ascorbate-glutathione pathway. GR activity significantly decreased at 10, 20, 30, 40, and 50 µg mL⁻¹ Cyp applications. Sáenz et al. (2012) found that Cyp concentrations causing algicidal effects, and thereby inhibitory effects of GR enzyme activity due to oxidative

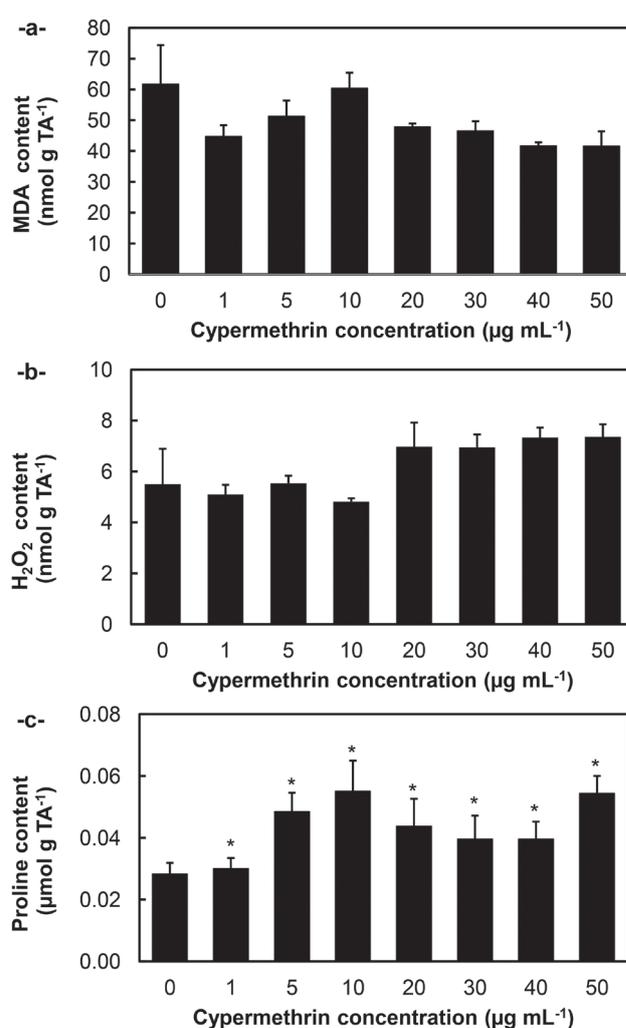


Fig. 4. Malondialdehyde (MDA; a), hydrogen peroxide (H₂O₂; b) and proline (c) contents of *Arthrospira platensis* treated with 0–50 µg mL⁻¹ cypermethrin. Data are the means ± SE of three replicates. *Significantly different from control, $P < 0.05$ (LSD analysis).

stress damage on *P. subcapitata*. The reductions in GR enzyme activities may have been related to loss of cell viability or deterioration of enzyme structure and enzyme reactions.

In contrast, GR activity did not change in Dlm application in our study. Dewez et al. (2005) treated a fungicide Fludioxonil to *Scenedesmus obliquus* and indicated that GR activity was not significantly affected and the oxidized glutathione pool may have also been used by other enzymes.

APX enzyme activity statistically increased at all concentrations in Cyp treatment. APX use ascorbic acid for the elimination of detrimental H₂O₂ (Verma and Dubey 2003). Studies on plants showed that APX activity increased in various stress conditions (Hideg and Vass 1996, Verma and Dubey 2003).

Teisseire and Vernet (2001) supported the conclusion that GR enzyme activity was related to the APX enzyme activity. In our study, it was concluded that the absence of alteration in GR activity at the application concentrations of Dlm promoted the absence of alteration in APX enzyme at

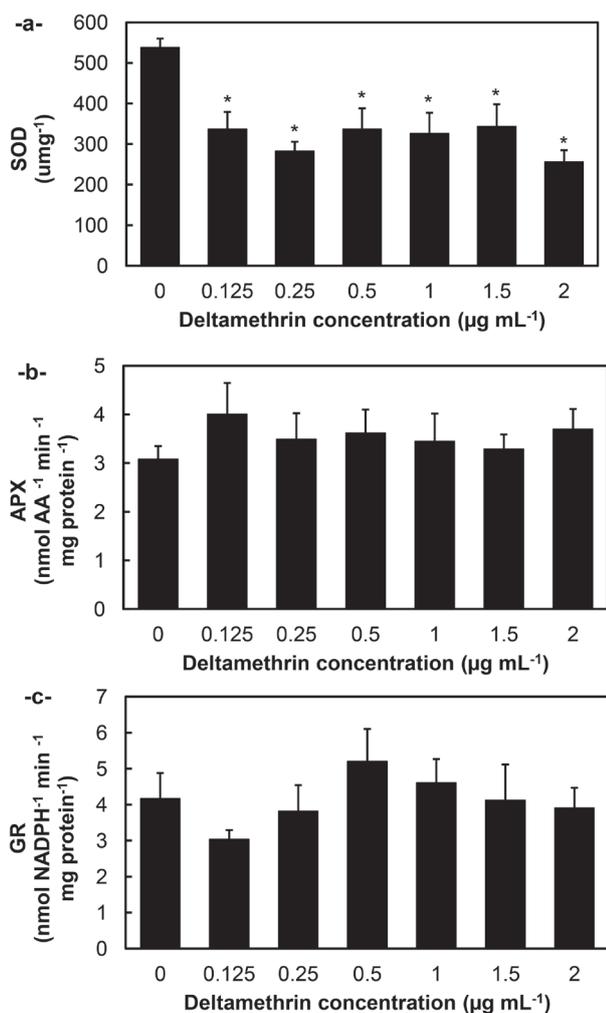


Fig. 5. Total superoxide dismutase (SOD; a), ascorbate peroxidase (APX; b) and glutathione reductase (GR; c) activities of *Arthrospira platensis* treated with 0–2 µg mL⁻¹ deltamethrin. Data are the means ± SE of three replicates. *Significantly different from control, $P < 0.05$ (LSD analysis).

similar concentrations because the ascorbate pool was counterbalanced by the GR enzyme.

In our study, MDA content in *A. platensis* cultures treated with Dlm significantly increased at 0.5 µg mL⁻¹, 1 µg mL⁻¹, and 1.5 µg mL⁻¹ concentrations. The changes in MDA content were parallel with the changes of H₂O₂ content for Dlm pesticide. The increase of H₂O₂ content caused the formation of OH radicals by the Haber-Weis reaction and therefore lipid peroxidation was increased (Goel and Sheoran 2003). Moreover, the inhibition of SOD activity caused accumulation of O₂⁻ in the medium in deltamethrin application. It is known that lipid peroxidation is linked to the amount of O₂⁻ (Choudhary et al. 2007). Also, Baruah and Chaurasia (2020) reported that alpha-cypermethrin decreased the photosynthetic pigment content and it increased the MDA content of *Chlorella* sp. at 6–48 mg L⁻¹ during 96 h bioassay. The reduction of photosynthetic pigments also increases lipid peroxidation (Chen et al. 2008). A significant reduction of chlorophyll-*a* content may take place with lipid peroxidation at Dlm application. It was reported that MDA content, an indicator of lipid peroxidation, was increased by the endosul-

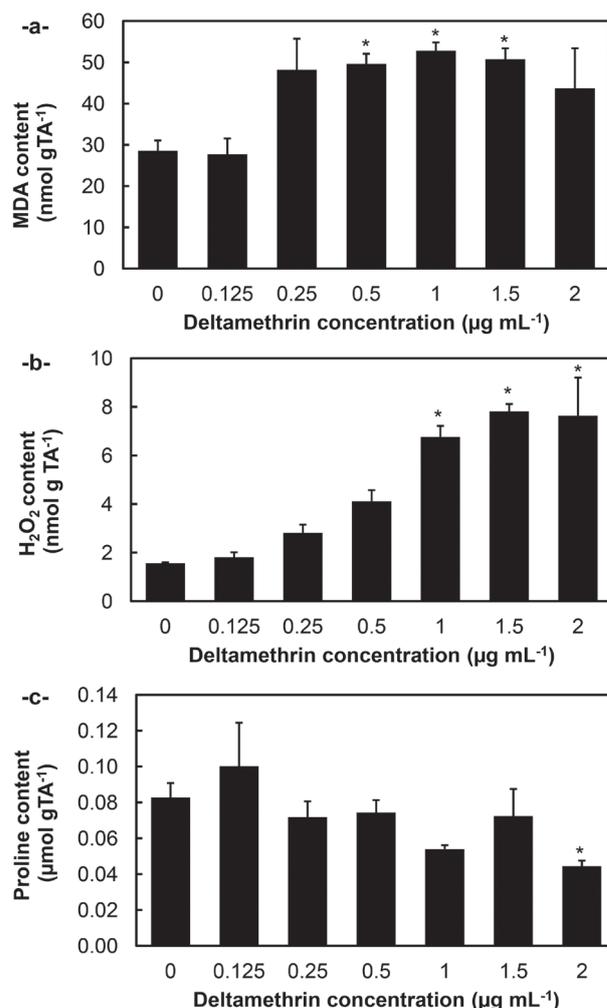


Fig. 6. Malondialdehyde (MDA; a), hydrogen peroxide (H₂O₂; b) and proline (c) contents of *Arthrospira platensis* treated with 0–2 µg mL⁻¹ deltamethrin. Data are the means ± SE of three replicates. *Significantly different from control, $P < 0.05$ (LSD analysis).

fan application (Kumar et al. 2008). Wang et al. (2011) found that cypermethrin enhanced MDA content on *Skeletonema costatum*, *Scripsiella trochoidea* and *Chatonella marina*.

There was no significant change in the MDA content at the Cyp application and it was supported by the results of H₂O₂ content. Moreover, the increase in the proline content may have prevented the cell membrane damage and the amount of MDA. Siripornadulsil et al. (2002) reported that the cadmium application did not alter the amount of MDA in the proline overproducing transgenic *Chlamydomonas reinhardtii* P.A.Dangeard, and they suggested that proline may have inhibited free radical damage due to its antioxidant action.

The amount of H₂O₂ in *A. platensis* cultures exposed to the Dlm significantly increase with addition of 1 µg mL⁻¹, 1.5 µg mL⁻¹, and 2 µg mL⁻¹ Dlm. It was expected that the application of Dlm pesticide would result in a decrease in the H₂O₂ content with the increase of SOD enzyme activity, but the H₂O₂ content likely increased due to the increased activity of oxidases such as glycolate oxidase, glucose oxidase, amino acid oxidase and sulfite oxidase in plants (Asada 1999).

The amount of H₂O₂ in all concentrations of cultures of *A. platensis* exposed to Cyp (except 1 µg mL⁻¹ and 5 µg mL⁻¹) was not changed by increased APX enzyme activity. The decrease in the amount of H₂O₂ is likely due to the increase in enzymes consuming the H₂O₂ content, such as the APX enzyme (Mallick and Mohn 2000).

The free proline amount of *A. platensis* cultures exposed to Cyp was statistically significantly increased at all concentrations compared to the control. Proline accumulation has been reported in plants exposed to heavy metal stress (Saradhi and Saradhi 1991). Proline is an effective singlet oxygen scavenger and regulates the cell redox potential (Saradhi and Saradhi 1991, Alia et al. 2001). Proline acts as an osmotic regulator and scavenger of OH radical in cells and thus it interacts with cell macromolecules such as DNA, protein, and lipids and consequently stabilizes the structure and function of these molecules (Kavir et al. 2005). The increase of proline content may be an adaptive response under extreme stress conditions (Fatma et al. 2007, Kumar et al. 2014). Proline reduces free radical production under stress conditions (Alia 1993). Fatma et al. (2007) studied the effects of environmental pollution by evaluating proline content on *Westiellopsis prolifica* Janet cyanobacterium and found that alphamethrin pesticide and heavy metals increased proline accumulation.

The free proline content of *A. platensis* cultures exposed to the Dlm effect was significantly reduced at 2 µg mL⁻¹ concentration. Ewald and Schlee (1983) found that sulfide decreased the proline content by inhibiting proline synthesis. Likewise, Dlm may have inhibited the proline synthesis or may have disrupted proline configuration in our Dlm application.

In conclusion, in this study, the decrease in biomass and chlorophyll-a was related to the concentration of Cyp and Dlm. Antioxidant enzyme activities and parameters were affected by different degrees according to the particular pesticides and their concentrations. These differences arose from the ROS producing capacity of Cyp and Dlm. Also, deltamethrin is more toxic than cypermethrin according to concentrations. Since these pesticides have a toxic effect on aquatic organisms, care should be taken when they are used with all necessary precautions to prevent their release into the aquatic ecosystem.

Acknowledgement

This study was supported by Sakarya University Research Projects under Grant no. BAP 2017-02-20-001 and BAP 2016-02-20-002.

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